3.D - Agricultural Soils

# 3.D - Agricultural Soils

## **Short description**

NFR-Code	Name of Category	Method	AD	EF	State of reporting
3.D	Agricultural Soils				
consisting	of / including source categories				
3.D.a.1	Inorganic N-fertilizers (includes also urea application)	T2 (NH <sub>3</sub> ), T1 (for NO <sub>x</sub> )	NS,RS	D (NH <sub>3</sub> ), D (NO <sub>x</sub> )	
3.D.a.2.a	Animal manure applied to soils	T2, T3 (NH <sub>3</sub> ), T1 (for NO <sub>x</sub> )	М	CS (NH <sub>3</sub> ), D (NO <sub>x</sub> )	
3.D.a.2.b	Sewage sludge applied to soils	T1 (for NH <sub>3</sub> ,NO <sub>x</sub> )	NS, RS	D (NH <sub>3</sub> ), D (NO <sub>x</sub> )	
311276	Other organic fertilisers applied to soils (including compost)	T2 (for NO <sub>x</sub> , NH <sub>3</sub> )	М	CS	
3.D.a.3	Urine and dung deposited by grazing animals	T1 (for NH <sub>3</sub> , NO <sub>x</sub> )	NS,RS	D	
3.D.c	Farm-level agricultural operations including storage, handling and transport of agricultural products	T1 (for TSP, PM <sub>10</sub> , PM <sub>2.5</sub> )	NS, RS	D	
3.D.d	Off-farm storage, handling and transport of bulk agricultural products				NA & for Black Carbon, NR
3.D.e	Cultivated crops	T2 (NMVOC)	NS, RS	D	
3.D.f	Agriculture other including use of pesticides	T2 (HCB)	NS	D	

<b>Key Category</b>	NO <sub>x</sub>	NMVOC	SO <sub>2</sub>	NΗ₃	PM <sub>2.5</sub>	PM <sub>10</sub>	TSP	ВС	co	Pb	Cd	Hg	Diox	PAH	нсв
3.D.a.1	L/-	-	-	L/T	-	-	-	-	-	-	-	-	-	-	-
3.D.a.2.a	L/-	-	-	L/T	-	-	-	-	-	-	-	-	-	-	-
3.D.a.2.b	-/-	-	-	-/-	-	-	-	-	-	-	-	-	-	-	-
3.D.a.2.c	-/-	-	-	L/T	-	-	-	-	-	-	-	-	-	-	-
3.D.a.3	-/-	-	-	-/-	-	-	-	-	-	-	-	-	-	-	-
3.D.c	-	-	-	-	-/-	L/-	L/-	-	-	-	-	-	-	-	-
3.D.e	-	-/-	-	-	-	-	-	-	-	-	-	-	-	-	-
3.D.f	-	-	-	-	-	-	-	-	-	-	-	-	-	-	L/-

T = key source by Trend L = key source by Level

Methods	
D	Default
RA	Reference Approach
T1	Tier 1 / Simple Methodology *
T2	Tier 2*
Т3	Tier 3 / Detailed Methodology *
С	CORINAIR
CS	Country Specific
М	Model
*	ADINAID Encircian Inventory Childhealt 2007 in the array and if a chartery

\* as described in the EMEP/CORINAIR Emission Inventory Guidebook - 2007, in the group specific chapters.

ΑD	- Data Source for Activity Data
NS	National Statistics
RS	Regional Statistics
IS	International Statistics
PS	Plant Specific data
AS	Associations, business organisations
0	specific guestionnaires, surveys

EF	- Emission Factors
D	Default (EMEP Guidebook)
С	Confidential

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#### **EF - Emission Factors**

**CS** Country Specific

**PS** Plant Specific data

### **Country specifics**



#### NH<sub>3</sub> and NO<sub>x</sub>

In 2020, agricultural soils emitted 260.0 kt  $NH_3$  or 50.7 % of the total agricultural  $NH_3$  emissions in Germany (512.3 kt  $NH_3$ ). The main contributions to the total  $NH_3$  emissions from agricultural soils are the application of manure (3.D.a.2.a), with 158.7 kt (61.0 %) and the application of inorganic N-fertilizers (3.D.a.1) with 50.1 kt (19.3 %).

Application of sewage sludge (3.D.a.2.b) contributes 0.7 % or 1.9 kt NH<sub>3</sub>.

The application of residues from the digestion of energy crops (3.D.a.2.c) leads to 37.0 kt  $NH_3$  or 14.2 %. N excretions on pastures (3.D.a.3) have a share of 12.3 kt  $NH_3$  or 4.7 %.

In 2020, agricultural soils were the source of 98.6 % (107.2 kt) of the total of  $NO_x$  emissions in the agricultural category (108.7 kt). The  $NO_x$  emissions from agricultural soils are primarily due to application of inorganic fertilizer (3.D.a.1) (50.1 %) and manure (3.D.a.2.a) (34.3 %). Application of residues from digested energy crops (3.D.a.2.c) contributes 10.8 % to agricultural soil emissions, 4.3 % are due to excretions on pastures (3.D.a.3). Emissions from application of sewage sludge (3.D.a.2.b) contribute 0.5 %.

#### **NMVOC**

In 2020, the category of agricultural soils contributed 9.2 kt NMVOC or 3.1 % to the total agricultural NMVOC emissions in Germany. The only emission source was cultivated crops (3.D.e).

#### TSP, PM<sub>10</sub> & PM<sub>2.5</sub>

In 2020, agricultural soils contributed, respectively, 28.7 % (17.3 kt), 57.2 % (17.3 kt) and 15.1 % (0.7 kt) to the total agricultural TSP,  $PM_{10}$  and  $PM_{2.5}$  emissions (60.3 kt, 30.2 kt, 4.4 kt, respectively). The emissions are reported in category 3.D.c (Farm-level agricultural operations including storage, handling and transport of agricultural products).

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## 3.D.a.1 - Inorganic N-fertilizers

The calculation of  $NH_3$  and NOx (NO) emissions from the application of inorganic fertilizers is described in Vos et al. (2022), Chapter 11.1  $^{1)}$ .

#### **Activity Data**

German statistics report the amounts of fertilizers sold which are assumed to equal the amounts that are applied. Since the 2021 submission, storage effects are approximated by applying a moving average to the sales data (moving centered three-year average, for the last year a weighted two-year average, which assigns 2/3 of the weight to the last year).

Table 1: AD for the estimation of NH<sub>3</sub> and NOx emissions from application of inorganic fertilizers

	Application of inorganic fertilizers in Gg N														
1990 1995 2000 2005 2010 2011 2012 2013 2014 2015 2016 2017 20											2018	2019	2020		
Application of fertilizers (total)	2'196	1'723	1'922	1'797	1'635	1'665	1'692	1'655	1'716	1'736	1'731	1'622	1'499	1'404	1'362
Calcium ammonium nitrate	1'368	1'044	982	824	689	708	680	644	633	618	605	571	543	520	508
Nitrogen solutions (urea AN)	127	223	261	236	180	187	181	173	173	172	171	162	151	137	132
Urea	243	180	247	290	362	323	348	342	391	417	433	377	310	248	225
Ammonium phosphates	85	55	66	55	64	71	77	78	82	84	82	77	65	64	63
Other NK and NPK	246	162	175	126	63	66	73	71	72	67	62	54	52	51	52
Other straight fertilizers	127	60	191	266	277	311	331	348	365	377	377	381	378	383	383

#### Methodology

 $NH_3$  emissions from the application of inorganic fertilizers are calculated using the Tier 2 approach according to EMEP (2019)-3D-14ff <sup>2)</sup>, distinguishing between various fertilizer types, see Table 2. For  $NO_x$ , the Tier 1 approach described in EMEP (2019) [10]-3D-11 is applied.

#### **Emission factors**

The emission factors for  $NH_3$  depend on fertilizer type, see EMEP (2019)-3D-15. Table 2 lists the EMEP emission factors for the fertilizers used in the inventory. In order to reflect average German conditions the emission factors for cool climate and a pH value lower than 7 was chosen. For urea fertilizer the German fertilizer ordinance prescribes the use of urease inhibitors or the immediate incorporation into the soil from 2020 onwards. The NH3 emission factor for urea fertilizers is therefore reduced by 70% from 2020 onwards, according to Bittman et al. (2014, Table 15) $^{3}$ ).

Table 2: NH<sub>3</sub>-EF for inorganic fertilizers

Inorganic fertilizers, emission factors in $kg\ NH_3$ per $kg\ fertilizer\ N$											
Fertilizer type	EF										
Calcium ammonium nitrate	0.008										
Nitrogen solutions (UREA AN)	0.098										
Urea	0.155 (in 2020: 0.0465)										
Ammonium phosphates	0.050										
Other NK and NPK	0.050										
Other straight fertilizers	0.010										

For  $NO_x$ , the simpler methodology by EMEP (2019)-3D-11 was used. The emission factor 0.040 from EMEP, 2019-3D, Table 3.1 has the units of kg  $N_2O$  per kg fertilizer N and was derived from Stehfest and Bouwman (2006)<sup>4)</sup>. The German inventory uses the emission factor 0.012 kg NO-N per kg N derived from Stehfest and Bouwman (2006). This is equivalent to an emission factor of 0.03943 kg  $NO_x$  per kg fertilizer N (obtained by multiplying 0.012 kg NO-N per kg N with the molar weight ratio 46/14 for  $NO_2$ : NO). The inventory uses the unrounded emission factor.

Table 3: Emission factor for NO<sub>x</sub> emissions from fertilizer application

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<b>Emission factor</b>	kg NO-N	per kg fertilizer N	kg $\mathrm{NO}_{\mathrm{x}}$ per kg fertilizer N
EF <sub>fert</sub>		0.012	0.039

#### **Trend discussion for Key Sources**

In the last years (and up to 2019 in dramatic fashion) fertilizer sales have decreased. Emissions have fallen accordingly. This is even more pronounced for  $NH_3$  than for  $NO_x$ , as total  $NH_3$  from the application of mineral fertilizers is, until the year 2019, very strongly correlated with the amount of urea applied (R2 = 0.89), the sales of which have decreased more than for all other mineral fertilizers. Since 2020 the negative trend is reinforced as urea fertilizer have to be either used with urease inhibitors or have to be incorporated into the soil directly, which causes 70% lower emissions (Bittman et al. 2014).

#### Recalculations

Table REC-1 shows the effects of recalculations on  $NH_3$  and  $NO_x$  emissions. The only differences are in 2019 as the year 2020 is now included in the weighted average.

Table REC-1: Comparison of NH<sub>3</sub> and NO<sub>x</sub> emissions from fertilizer application of the submissions (SUB) 2021 and 2022

			N	H₃ and	d NO <sub>x</sub>	emiss	ions f	rom f	ertiliz	er apı	olicati	on, in	Gg			
	SUB 1990 1995 2000 2005 2010 2011 2012 2013 2014 2015 2016 2017 2018 2019 2020															
NH <sub>3</sub>	2022	78.82	69.56	85.64	86.36	88.43	83.96	88.04	85.95	93.92	97.89	99.73	89.25	76.79	65.63	36.97
NH <sub>3</sub>	2021	78.82	69.56	85.64	86.36	88.43	83.96	88.04	85.95	93.92	97.89	99.73	89.25	76.79	68.09	
NO <sub>x</sub>	2022	86.57	67.94	75.77	70.84	64.48	65.66	66.71	65.25	67.65	68.46	68.24	63.95	59.11	55.34	53.71
NO <sub>x</sub>	2021	86.57	67.94	75.77	70.84	64.48	65.66	66.71	65.25	67.65	68.46	68.24	63.95	59.11	55.97	

#### **Planned improvements**

No improvements are planned at present.

## 3.D.a.2.a - Animal manure applied to soils

In this sub category Germany reports the  $NH_3$  and  $NO_x$  (NO) emissions from application of manure (including application of anaerobically digested manure). For an overview see Vos et al. (2022), Chapter 11.2.

#### **Activity data**

The calculation of the amount of N in manure applied is based on the N mass flow approach (see 3.B). It is the total of N excreted by animals in the housing and the N imported with bedding material minus N losses by emissions of N species from housing and storage. Hence, the amount of total N includes the N contained in anaerobically digested manures to be applied to the field.

The frequencies of application techniques and incorporation times as well as the underlying data sources are described in Vos et al. (2022), Chapter 3.4.3. The frequencies are provided e. g. in the NIR 2022<sup>5)</sup>, Chapter 19.3.2.

Table 4: AD for the estimation of  $NH_3$  and  $NO_x$  emissions from application of manure

		Application of manure in Gg N  990 1995 2000 2005 2010 2011 2012 2013 2014 2015 2016 2017 2018 2019 2020														
199	90	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020	
1'1	20	972	954	924	928	933	949	961	972	972	966	961	947	940	932	

#### Methodology

 $NH_3$  emissions from manure application are calculated separately for each animal species in the mass flow approach by multiplying the respective TAN amount with  $NH_3$  emission factors for the various manure application techniques. For details see [3-b-manure-management 3.B] and Vos et al. (2022), Chapter 4 to 8 and 11.3. For  $NO_x$  emissions from manure

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application the inventory calculates NO-N emissions (see Vos et al. (2022), Chapter 11.2, that are subsequently converted into  $NO_x$  emissions by multiplying with the molar weight ratio 46/14. The Tier 1 approach for the application of inorganic fertilizer as described in EMEP (2019)-3D-11 is used, as no specific methodology is available for manure application.

#### **Emission factors**

Table 5 shows the time series of the overall German NH<sub>3</sub> IEF defined as the ratio of total NH<sub>3</sub>-N emission from manure application to the total amount of N spread with manure.

Table 5: IEF for NH₃-N from application of manure

	IEF in kg NH <sub>3</sub> -N per kg N in applied manure														
199	0	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
0.20	)2 (	0.187	0.180	0.168	0.162	0.162	0.157	0.155	0.152	0.150	0.148	0.147	0.145	0.143	0.140

For NO<sub>x</sub> the same emission factor as for the application of inorganic fertilizer was used (see Table 3).

#### **Trend discussion for Key Sources**

Both  $NH_3$  and  $NO_x$  emissions from the application of animal manures are key sources. Total  $NO_x$  is calculated proportionally to the total N in the manures applied which decreased remarkably from 1990 to 1991 due to the decline in animal numbers following the German reunification (reduction of livestock numbers in Eastern Germany). In the 1990s and 2000s this was followed by a weakened decline in animal manure amounts. From 2010 to 2014 there was a slight increase and since then the amount of N in manure applied has been slightly declining again, see Table 4. The  $NO_x$  emissions follow these trends. For total  $NH_3$  emissions there is a slight negative trend. This is due to the increasing use of application practices with lower  $NH_3$  emission factors.

#### Recalculations

Table REC-2 shows the effects of recalculations on  $NH_3$  and  $NO_x$ . The total emissions of  $NH_3$  and  $NO_x$  from application of manure are slightly lower than those of last year's submission from the year 2000 onwards for  $NH_3$ , from the year 2010 onwards for  $NO_x$ . In earlier years the emissions are slightly higher than in last year's submission.

These differences are predominantly caused by the update of data from the official agricultural census 2020 as well as the update of the suckler-cow model and the new raw protein contents in feed of fattening pigs and broilers, see main page of the agricultural sector, list of **recalculation reasons**, **No. 1**, **4**, **7 and 8**.

Much smaller is the impact of the updates of activity data for male cattle > 2 years, pigs, poultry and sheep (see **recalculation reasons 5, 6 and 9 through 12**) Further details on recalculations are described in Vos et al. (2022), Chapter 3.5.2.

Table REC-2: Comparison of the NH₃ and NO₂ emissions of the submissions (SUB) 2021 and 2022

	NH₃ and NO₂ emissions from application of manure, in Gg															
	SUB	1990	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
NH <sub>3</sub>	2022	275.21	221.15	208.05	188.31	182.09	183.07	180.74	181.30	179.97	177.25	174.11	171.06	166.32	162.64	158.67
NH <sub>3</sub>	2021	273.67	220.82	208.69	190.07	185.28	186.32	184.07	184.62	183.26	180.08	179.11	178.15	175.65	174.11	
NO <sub>x</sub>	2022	44.14	38.33	37.61	36.42	36.58	36.81	37.43	37.88	38.34	38.31	38.07	37.91	37.35	37.05	36.76
NO <sub>x</sub>	2021	43.46	37.99	37.41	36.35	36.71	36.99	37.67	38.18	38.70	38.58	38.39	38.27	37.80	37.54	

#### **Planned improvements**

No improvements are planned at present.

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## 3.D.a.2.b - Sewage sludge applied to soils

The calculation of  $NH_3$  and  $NO_x$  (NO) emissions from application of sewage sludge is described in Vos et al. (2022), Chapter 11.4.

#### **Activity data**

N quantities from application of sewage sludge were calculated from data of the German Environment Agency and (since 2009) from data of the Federal Statistical Office (see Table 6).

Table 6: AD for the estimation of NH<sub>3</sub> and NO<sub>x</sub> emissions from application of sewage sludge

1	Appli	catior	of se	ewage	slud	ge in	Gg N								
1	L990	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
Г	27	35	33	27	26	25	25	22	21	19	19	14	12	14	14

#### Methodology

A tier 1 methodology is used (EMEP, 2019, 3D, Chapter 3.3.1).  $NH_3$  and  $NO_x$  emissions are calculated by multiplying the amounts of N in sewage sludge applied with the respective emission factors.

#### **Emission factors**

EMEP (2019)-3.D, Table 3-1 provides a Tier 1 emission factor for  $NH_3$  (0.13 kg NH3 per kg N applied) emissions from application of sewage sludge. The German inventory uses the equivalent emission factor in  $NH_3$ -N units which is 0.11 kg  $NH_3$ -N per kg N applied (cf. the derivation of the emission factor described in the appendix of EMEP (2019)-3D, page 26-27). For  $NO_v$  the same emission factor like for the application of inorganic fertilizer was used (see Table 3).

#### **Trend discussion for Key Sources**

NH<sub>3</sub> and NO<sub>x</sub> emissions from the application of sewage sludge are no key sources.

#### Recalculations

Table REC-3 shows the effects of recalculations on  $NH_3$  and  $NO_x$  emissions. The only change compared to last year's submission occurs for the year 2018 and 2019 due to the update of the activity data (see main page of the agricultural sector, **recalculation No 15**. Further details on recalculations are described in Vos et al. (2022), Chapter 3.5.2.

Table REC-3: Comparison of the NH<sub>3</sub> and NO<sub>x</sub> emissions of the submissions (SUB) 2021 and 2022

NH <sub>3</sub>	and l	NO <sub>x</sub> eı	missio	ons fr	om ap	plicat	ion o	f sewa	age sl	udge,	in Gg	ı				
	SUB	1990	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
NH <sub>3</sub>	2022	3.66	4.71	4.40	3.66	3.48	3.35	3.33	2.87	2.85	2.50	2.50	1.89	1.67	1.90	1.90
NH <sub>3</sub>	2021	3.66	4.71	4.40	3.66	3.48	3.35	3.33	2.87	2.85	2.50	2.50	1.89	1.73	1.73	
NO <sub>x</sub>	2022	1.08	1.39	1.30	1.08	1.03	0.99	0.98	0.85	0.84	0.74	0.74	0.56	0.49	0.56	0.56
NO <sub>x</sub>	2021	1.08	1.39	1.30	1.08	1.03	0.99	0.98	0.85	0.84	0.74	0.74	0.56	0.51	0.51	

#### **Planned improvements**

No improvements are planned at present.

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## 3.D.a.2.c - Other organic fertilizers applied to soils

This sub category describes Germany's  $NH_3$  and  $NO_x$  (NO) emissions from application of residues from digested energy crops. For details see Vos et al. (2022), Chapters 10.2 and 11.3.

#### **Activity data**

Activity data is the amount of N in residues from anaerobic digestion of energy crops when leaving storage. This amount of N is the N contained in the energy crops when being fed into the digestion process minus the N losses by emissions of N species from the storage of the residues (see 3.I). N losses from pre-storage are negligible and there are no N losses from fermenter (see Vos et al. (2022), Chapter 10.2.1).

Table 7: AD for the estimation of  $NH_3$  and  $NO_x$  emissions from application of residues from anaerobic digestion of energy crops

Appli	catior	of re	sidue	s from	digest	ed ene	rgy pla	nts in (	Gg N					
1990	990 1995 2000 2005 2010 2011 2012 2013 2014 2015 2016 2017 2018 2019 2020													
0.05	0.62	5.40	45.76	167.41	209.32	230.52	279.13	292.42	303.81	302.16	297.19	292.86	293.08	293.08

#### Methodology

The  $NH_3$  emissions are calculated the same way as the  $NH_3$  emissions from application of animal manure (3.D.a.2.a). The frequencies of application techniques and incorporation times as well as the underlying data sources are provided e. g. in the NIR 2022, Chapter 19.3.2. The amounts of TAN in the residues applied are obtained from the calculations of emissions from the storage of the digested energy crops (3.I).

For  $NO_x$  emissions from application of residues the Tier 1 approach for the application of inorganic fertilizer as described in EMEP (2019)-3D-11 is used. The inventory calculates NO emissions that are subsequently converted into  $NO_x$  emissions by multiplying with the molar weight ratio 46/30.

#### **Emission factors**

For  $NH_3$  the emission factors for untreated cattle slurry were adopted, see Vos et al. (2022), Chapter 10.2. As the  $NO_x$  method for fertilizer application is used for the calculation of  $NO_x$  emissions from the application of residues, the emission factor for fertilizer application was used (see Vos et al. (2022), Chapter 11.1)

Table 8 shows the implied emission factors for  $NH_3$  emissions from application of residues from digested energy crops.

Table 8: IEF for NH<sub>3</sub>-N

IEF in	kg N	H₃-N p	er kg	N in o	digest	ed en	ergy o	rops						
1990	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
0.182	0.182	0.183	0.183	0.183	0.184	0.174	0.166	0.159	0.153	0.150	0.148	0.146	0.143	0.141

#### **Trend discussion for Key Sources**

The application of residues from anaerobic digestion of energy crops is a key source for  $NH_3$ . Emissions are dominated by the amounts of N in the substrates fed into the digestion process and to a lesser extent by the increased use of application techniques with lower emission factors. They have become important since about 2005 and have risen sharply until 2013. Since then, they have changed little each year and tend to decrease slightly in the last few years. The latter is mostly due to a small negative trend of the amounts of energy crops digested.

#### Recalculations

Table REC-4 shows the effects of recalculations on  $NH_3$  and  $NO_x$  emissions. The only changes compared to last year's submission occur for the years 2015-2019, due to the use of new data on manure spreading techniques from the official

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agricultural census 2020 (see main page of the agricultural sector, list of **recalculation reasons, No 1 and 16**, and Vos et al. (2022), Chapter 3.5.2.)

Table REC-4: Comparison of the NH3 and NOx emissions of the submissions (SUB) 2021 and 2022

NH <sub>3</sub>	and l	NO <sub>x</sub> eı	missic	ons fr	от ар	plicat	ion of	diges	sted e	nergy	crops	s, in G	g			
	SUB	1990	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
NΗ₃	2022	0.01	0.14	1.20	10.15	37.27	46.75	48.81	56.27	56.56	56.42	55.16	53.47	51.82	50.98	50.12
NH <sub>3</sub>	2021	0.01	0.14	1.20	10.15	37.27	46.75	48.81	56.27	56.56	56.42	56.11	55.37	54.63	54.63	
NO <sub>x</sub>	2022	0.00	0.02	0.21	1.80	6.60	8.25	9.09	11.01	11.53	11.98	11.91	11.72	11.55	11.56	11.56
NO <sub>x</sub>	2021	0.00	0.02	0.21	1.80	6.60	8.25	9.09	11.01	11.53	11.98	11.91	11.72	11.55	11.55	

#### **Planned improvements**

No improvements are planned at present.

## 3.D.a.3 - Urine and dung deposited by grazing animals

The calculation of NH<sub>3</sub> and NO<sub>x</sub> (NO) emissions from N excretions on pasture is described in Vos et al. (2022), Chapter 11.5.

#### **Activity data**

Activity data for  $NH_3$  emissions during grazing is the amount of TAN excreted on pasture while for  $NO_x$  emissions it is the amount of N excreted on pasture.

Table 9 shows the N excretions on pasture. The TAN excretions are derived by multiplying the N excretions with the relative TAN contents provided in 3.B, Table 2.

Table 9: N excretions on pasture

N excretion	ns on	pastu	re in '	% of t	otal N	l excr	eted								
	1990	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
Dairy cows	20.3	15.6	12.7	11.3	10.3	10.1	9.8	9.5	9.2	8.9	8.6	8.3	8.0	7.7	7.4
Other cattle	15.1	17.3	18.9	19.0	19.4	19.5	19.6	19.7	19.9	20.3	20.5	20.7	20.9	21.2	21.5
Sheep	55.1	55.5	55.1	55.4	54.8	55.1	55.1	55.2	55.3	55.4	55.4	55.4	55.6	55.5	55.4
Goats	34.2	34.2	34.2	34.2	34.2	34.2	34.2	34.2	34.2	34.2	34.2	34.2	34.2	34.2	34.2
Horses	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5	20.5

#### Methodology

 $NH_3$  emissions from grazing are calculated by multiplying the respective animal population (3.B, Table 1) with corresponding N excretions and relative TAN contents (3.B, Table 2) and the fraction of N excreted on pasture (Table 9). The result is multiplied with the animal specific emission factor (Table 10). NO emissions are calculated the same way with the exception that the emission factor is related to N excreted instead of TAN.

#### **Emission Factors**

The emission factors for  $NH_3$  are taken from EMEP (2019)-3B-31, Table 3.9. They relate to the amount of TAN excreted on pasture. Following the intention of EMEP, 2019-3D, Table 3.1, the inventory uses for  $NO_x$  the same emission factor as for the application of inorganic fertilizer (see Table 3). In order to obtain  $NO_x$  emissions (as  $NO_2$ ) the NO-N emission factor of 0.12 kg NO-N per kg N excreted is multiplied by 46/14.

Table 10: Emission factors for emissions of NH₃ and NO from grazing

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<b>Emission factors</b>	
Dairy cows	0.14 kg NH3-N per kg TAN excreted
Other cattle	0.14 kg NH3-N per kg TAN excreted
Horses	0.35 kg NH3-N per kg TAN excreted
Sheep, goats	0.09 kg NH3-N per kg TAN excreted
All animals	0.012 kg NO-N per kg N excreted

#### **Trend discussion for Key Sources**

Emissions from urine and dung deposited by grazing animals are no key sources.

#### Recalculations

Table REC-5 shows the effects of recalculations on NH<sub>3</sub> and NOx emissions.

Because overall N excretions on pasture are lower than in last year's submission for dairy cattle, but higher for other cattle (predominantly due to new cattle grazing data from the official agricultural census 2020 and the update of the suckler-cow models, see main page of the agricultural sector, list of **recalculation reasons**, **No 1 and 2**),  $NO_x$  emissions are lower than in last year's submission. However, although  $NH_3$  emissions could be expected to show the same pattern, they are slightly higher than in the last submission in 1990-2013 and slightly lower than in last year's submission from 2014 onwards. This is due to a combination of two effects: The TAN content of the suckler cows' N-excretions is higher in the new suckler-cow model, which leads to higher emissions over the whole time series. This is over-compensated by declining dairy cow grazing times from 2014 onwards, which leads to lower emissions (see list of **recalculation reasons**, **No 1 and 2**). Further details on recalculations are described in Vos et al. (2022), Chapter 3.5.2.

Table REC-5: Comparison of the NH<sub>3</sub> and NO<sub>4</sub> emissions of the submissions (SUB) 2021 and 2022

NH <sub>3</sub>	and l	NO <sub>x</sub> er	nissio	ns fro	m gra	zing,	in Gg									
	SUB	1990	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
NH <sub>3</sub>	2022	22.23	18.15	16.26	14.35	13.80	13.43	13.29	13.37	13.40	13.40	13.20	13.03	12.74	12.56	12.30
NH <sub>3</sub>	2021	22.16	18.04	16.10	14.21	13.61	13.30	13.22	13.35	13.43	13.51	13.34	13.20	12.93	12.78	
NO <sub>x</sub>	2022	8.40	6.82	6.14	5.45	5.23	5.08	5.02	5.04	5.06	5.05	4.97	4.90	4.79	4.73	4.62
NO <sub>x</sub>	2021	8.44	6.89	6.22	5.53	5.30	5.17	5.15	5.20	5.25	5.29	5.24	5.20	5.13	5.10	

#### Planned improvements

No improvements are planned at present.

# 3.D.c - Farm-level agricultural operations including storage, handling and transport of agricultural products

In this category Germany reports TSP,  $PM_{10}$  and  $PM_{2.5}$  emissions from crop production according to EMEP (2019)-3D-11. For details see Vos et al. (2022), Chapter 11.14.

#### **Activity data**

The activity data is the total area of arable and horticultural land. This data is provided by official statistics.

Table 11: AD for the estimation of TSP,  $PM_{10}$  and  $PM_{2.5}$  emissions from soils

Arable	and h	orticult	tural la	nd in 1	000*ha	3								
1990	1990   1995   2000   2005   2010   2011   2012   2013   2014   2015   2016   2017   2018   2019   2020													
11'179	10'257	10'683	10'902	11'411	11'431	11'421	11'478	11'475	11'346	11'281	11'273	11'181	11'163	11'071

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#### Methodology

As the Tier 2 methodology described in EMEP (2019)-3D-17 cannot be used due to lack of input data, the Tier 1 methodology described in EMEP(2019)-3D-11 is used.

#### **Emission factors**

Emission factors given in EMEP (2019)-3D-12 are used. The Guidebook does not indicate whether EFs have considered the condensable component (with or without).

Table 12: Emission factors for PM emissions from agricultural soils

<b>Emission factor</b>	kg	ha <sup>-1</sup>
EF <sub>TSP</sub>	1.	56
EF <sub>PM10</sub>	1.	56
EF <sub>PM2.5</sub>	0.	06

#### **Trend discussion for Key Sources**

TSP and  $PM_{10}$  are key sources. Emissions depend only on the areas covered. These are relatively constant, with a very slight decrease over the past 10 years.

#### **Recalculations**

Table REC-6 shows the effects of recalculations on particulate matter emissions. There are minor changes with respect to last year's submission in several years because of updates of cultivation areas (see main page of the agricultural sector, list of **recalculation reasons, No 17**). However, due to the data format in Table REC-6, these differences are not visible. Further details on recalculations are described in Vos et al. (2022), Chapter 3.5.2.

Table REC-6: Comparison of particle emissions (TSP, PM<sub>10</sub> & PM<sub>2.5</sub>) of the submissions (SUB) 2021 and 2022

TSP,	PM <sub>10</sub> ,	PM <sub>2.5</sub>	emiss	sions 1	from o	rop p	roduc	tion, i	in Gg							
	SUB	1990	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
TSP	2022	17.44	16.00	16.67	17.01	17.80	17.83	17.82	17.91	17.90	17.70	17.60	17.59	17.44	17.41	17.27
TSP	2021	17.44	16.00	16.67	17.01	17.80	17.83	17.82	17.91	17.90	17.70	17.60	17.59	17.44	17.41	
PM <sub>10</sub>	2022	17.44	16.00	16.67	17.01	17.80	17.83	17.82	17.91	17.90	17.70	17.60	17.59	17.44	17.41	17.27
PM <sub>10</sub>	2021	17.44	16.00	16.67	17.01	17.80	17.83	17.82	17.91	17.90	17.70	17.60	17.59	17.44	17.41	
PM <sub>2.5</sub>	2022	0.67	0.62	0.64	0.65	0.68	0.69	0.69	0.69	0.69	0.68	0.68	0.68	0.67	0.67	0.66
PM <sub>2.5</sub>	2021	0.67	0.62	0.64	0.65	0.68	0.69	0.69	0.69	0.69	0.68	0.68	0.68	0.67	0.67	

#### **Planned improvements**

No improvements are planned at present.

## 3.D.e - Cultivated crops

In this category Germany reports NMVOC emissions from crop production according to EMEP (2019)-3D-16. For details see Vos et al. (2022), Chapter 11.12.

#### **Activity data**

The activity data is the total area of arable land and grassland. This data is provided by official statistics.

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Table 13: AD for the estimation of NMVOC emissions from crop production

Arable	land a	nd gra	ssland	in 100	0*ha									
1990	1990   1995   2000   2005   2010   2011   2012   2013   2014   2015   2016   2017   2018   2019   2020													
16'506	15'312	15'498	15'561	15'734	15'752	15'729	15'769	15'802	15'719	15'662	15'647	15'570	15'563	15'447

#### Methodology

The Tier 2 methodology described in EMEP (2019)-3D-16ff is used.

#### **Emission Factors**

The emission factors for wheat, rye, rape and grass (15°C) given in EMEP (2019)-3D-16, Table 3.3 were used. For all grassland areas the grass (15°C) EF is used, for all other crops except rye and rape the EF of wheat is used. Table 14 shows the implied emission factors for NMVOC emissions from crop production. The implied emission factor is defined as ratio of the total NMVOC emissions from cultivated crops to the total area given by activity data.

Table 14: IEF for NMVOC emissions from crop production

IE	IEF for NMVOC emissions from crop production in kg ha. <sub>1</sub>														
1	990	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
	0.47	0.53	0.57	0.59	0.61	0.57	0.64	0.66	0.72	0.63	0.62	0.62	0.50	0.55	0.59

#### **Trend discussion for Key Sources**

NMVOC emissions from crop production are no key sources.

#### Recalculations

Table REC-7 shows the effects of recalculations on NMVOC emissions. There are minor changes with respect to last year's submission in seral years because of updates of yields (see main page of the agricultural sector, list of **recalculation reasons, No 17**). However, due to the data format in Table Table REC-6, these differences are not visible. Further details on recalculations are described in Vos et al. (2022), Chapter 3.5.2.

Table REC-7: Comparison of NMVOC emissions of the submissions (SUB) 2021 and 2022

NMVOC emissions from crop production, in Gg															
SUB	1990	1995	2000	2005	2010	2011	2012	2013	2014	2015	2016	2017	2018	2019	2020
2022	7.69	8.19	8.79	9.17	9.53	9.03	10.05	10.36	11.40	9.91	9.69	9.74	7.82	8.56	9.16
2021	7.69	8.19	8.79	9.17	9.53	9.03	10.05	10.36	11.40	9.91	9.69	9.74	7.82	8.56	

#### **Planned improvements**

No improvements are planned at present.

#### Uncertainty

Details are described in chapter 1.7.

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EMEP (2019): EMEP/EEA air pollutant emission inventory guidebook – 2019, EEA Report No 13/2019, https://www.eea.europa.eu/publications/emep-eea-guidebook-2019.

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Stehfest E., Bouwman L. (2006): N2O and NO emission from agricultural fields and soils under natural vegetation: summarizing available measurement data and modelling of global emissions. Nutr. Cycl. Agroecosyst. 74, 207 – 228.

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